# **Peer Community Journal**

**Section: Ecology** 

# **RESEARCH ARTICLE**

**Published** 2023-09-12

# Cite as

Michael Raatz (2023) Provision of essential resources as a persistence strategy in food webs, Peer Community Journal, 3: e82.

> **Correspondence** mraatz@evolbio.mpg.de

# **Peer-review**

Peer reviewed and recommended by PCI Ecology, https://doi.org/10.24072/pci. ecology.100527

# (cc) BY

This article is licensed under the Creative Commons Attribution 4.0 License.

# Provision of essential resources as a persistence strategy in food webs

Michael Raatz<sup>0,1</sup>

Volume 3 (2023), article e82

https://doi.org/10.24072/pcjournal.315

# Abstract

Pairwise interactions in food webs, including those between predator and prey are often modulated by a third species. Such higher-order interactions are important structural components of natural food webs that can increase the stability of communities against perturbations and ensure continued ecosystem functioning. Particularly the flux of rare organic and inorganic compounds that are essential to species in the community can create higher-order interactions. Even though many such compounds exist, their effect on structuring communities is little understood. In this study, I perform invasion analyses on a general food web model that depicts apparent and exploitative competition. Introducing the provision of essential resources by a prey species to either its competitor or its predator as a higher-order interaction, I find that this mechanism can ensure the focal prey's persistence. Larger dietary essentiality, i.e. a stronger dependence of the predator or the competitor on the essential resource can increase the invasion growth rate of the focal prey to positive values, thus promoting its persistence when it would go extinct for low essentiality. This research shows that essential resources and the higher-order interactions created by them should be considered in community ecology.

<sup>1</sup>Department of Theoretical Biology, Max Planck Institute for Evolutionary Biology, Plön, Germany



Peer Community Journal is a member of the Centre Mersenne for Open Scientific Publishing http://www.centre-mersenne.org/

e-ISSN 2804-3871



#### Introduction

Growth, reproduction and survival of organisms can be limited by organic and inorganic compounds, which often are not present in the organism's diet in favourable concentrations or ratios. Particularly consumers at the plant-herbivore interface are often affected by dietary mismatches (Elser et al., 2016; Gaedke, Hochstädter, and Straile, 2002; Urabe, Shimizu, and Yamaguchi, 2018; Wacker and Martin-Creuzburg, 2012). This motivated considering besides food quantity also the quality of food when investigating performance measures of aquatic consumers (Andersen, Elser, and Hessen, 2004; Anderson and Hessen, 2005; Guo et al., 2016; Koussoroplis et al., 2019; Schälicke et al., 2019; Wacker and Martin-Creuzburg, 2012), terrestrial herbivores (Douglas, 2015; Eberl et al., 2020) and pollinators (Filipiak et al., 2017). The scarcity of resources that are essential for growth and reproduction but cannot be easily acquired from the environment can constrain the flow of matter and energy between trophic levels. Therefore, dietary limitations induced by essential resources can have important effects on population and community dynamics (Burian, Nielsen, and Winder, 2020; Gaedke, Hochstädter, and Straile, 2002; Iwabuchi and Urabe, 2012; Muller et al., 2001; Raatz, Gaedke, and Wacker, 2017; Schade et al., 2003; Singer et al., 2012; Stiefs et al., 2010).

Dietary dependencies also regularly exist within the same trophic level, where uptake of essential resources occurs from the environment, such as within the microbial loop when bacteria consume dissolved organic carbon from phytoplankton exudates (Azam et al., 1983; Pomeroy et al., 2007) or during the exchange of essential nutrients and metabolites between bacteria and microalgae (Croft et al., 2005; D'Souza et al., 2018; Kazamia et al., 2012; Oña and Kost, 2022; Soria-Dengg, Reissbrodt, and Horstmann, 2001; Suleiman et al., 2016). Understanding the mechanisms and effects of such dependencies is crucial for biodiversity research given for example the importance of microalgae such as diatoms for aquatic ecosystems and global carbon dynamics (Croft et al., 2005; Koedooder et al., 2019), but also for medical fields like human microbiome research (Herren, 2020) and antibiotic resistance research (Adamowicz et al., 2018). Taken together, dietary mismatches and dependencies of organisms from the same or different trophic levels are crucial determinants for the structure of their communities.

Mechanistically, community structure is determined by direct interactions within pairs of species or by indirect interactions across multiple species from the same or different trophic levels, e.g. through trophic cascades or apparent competition. Additionally to direct and indirect interactions, higher-order interactions, here defined as the density of a third species affecting the interaction of two other species (sensu Billick and Case, 1994), were found to potentially structure communities. The effects of higher-order interactions include stabilizing population dynamics (Grilli et al., 2017), increasing robustness against perturbation (Gibbs et al., 2023; Terry, Morris, and Bonsall, 2019), determining fitness of competitors (Mayfield and Stouffer, 2017) and affecting biodiversity-ecosystem-functioning relationships (Miele et al., 2019). Examples for higher-order interactions include trait-mediated effects such as a predator affecting the foraging rate of its prey or the prey's predation risk from other predators, and environment-mediated effects such as one species providing refuge to another species (Miele et al., 2019; Wootton, 2002).

In this paper, I will investigate another, so far unrecognized mechanism for creating higherorder interactions that arises from the provisioning of essential resources. In the presence of dietary mismatches one species, from here on referred to as the focal species, may provide resources that are essential to other community members. Such interactions are possible both towards members of the same trophic level, such as competitors, or towards members of different trophic levels, e.g. shared predators that prey on multiple species. For example, a higherorder interaction within the same trophic level is created when a competitor is co-limited by two resources but can only obtain one of those two resources from its environment and relies on another prey (the focal prey) to provide the other co-limiting resource. This provision may occur for example by leakage of common goods (Gore, Youk, and Van Oudenaarden, 2009) or carbon exudation in otherwise carbon-limited environments (Bratbak and Thingstad, 1985; Raatz, Schälicke, et al., 2018). A higher-order interaction between different trophic levels can arise when a predator obtains energy from multiple prey species but only the focal prey may provide additional, essential resources, e.g. vitamins or polyunsaturated fatty acids (Wacker and Martin-Creuzburg, 2012). Excess essential resources provided by the focal prey may then be used to efficiently convert other low-quality prey into predator biomass (Raatz, Gaedke, and Wacker, 2017).

In these two cases the provision of essential resources by the focal prey creates a higherorder interaction that manifests as an interaction modification (sensu Terry, Morris, and Bonsall, 2019) that regulates the flow of matter to the competitor or predator compartment in these communities, respectively. Regulating such fluxes has the potential to affect the biomass distribution in the community, ultimately determining the persistence of individual species. If such higher-order interactions increase the persistence of the focal prey they pose as an example for a niche-improving form of niche construction and they may thus even be adaptive (Kylafis and Loreau, 2008, 2011; Laland, Matthews, and Feldman, 2016). Consequently, in this paper, I will establish the provision of essential resources in a community as a mechanism driving higherorder interactions that may increase the persistence of the focal prey species and prevent its extinction either from predation or competition.

#### Methods

Investigating persistence of a focal species typically employs invasion analysis, which determines the net growth rate of that species in the remaining resident community when it is rare (and assumed to be invading) (Chesson, 1994; Ellner et al., 2019; MacArthur and Levins, 1967). If the focal prey provides the essential resources to some components of the community, being rare equates to switching off the higher-order interaction. Invasion analysis is therefore the perfect tool for determining the effect of essential resources both on the resident community and the persistence of the focal prey. Accordingly, I will investigate the invasion growth rate of the focal prey species  $X_1$  in a community that contains an abiotic resource R, a competing species  $X_2$  and a shared predator Y (Eqn. 1, Fig. 1), to incorporate essential resource provision in food webs. I assume a chemostat-type model in which the abiotic resource R is provided at a constant rate  $R_0 \delta$  and all entities experience the same dilution, see Tab. 1 for parameter definitions and values. The two prey species  $X_1$  and  $X_2$  take up the abiotic resource R at some rate r u(R), where u(R) defines the functional form of prey resource uptake. Both prey species are consumed by the predator following a functional response  $f(X_1, X_2)$ . I assume that the two prey species differ only in their attack probability (sensu Ehrlich and Gaedke, 2018) by a factor p and their maximum growth rate by a factor  $\phi$ . For example, p < 1 and  $\phi < 1$  implements a growth-defense trade-off (Fig. 2b).

(1)  

$$\frac{dR}{dt} = (R_0 - R) \delta - r u(R) X_1 - \mu_X(X_1) \phi r u(R) X_2$$

$$\frac{dX_1}{dt} = r u(R) X_1 - f(X_1, X_2) X_1 Y - \delta X_1$$

$$\frac{dX_2}{dt} = \varepsilon_X(X_1) \mu_X(X_1) \phi r u(R) X_2 - \mu_Y(X_1) \rho f(X_1, X_2) X_2 Y - \delta X_2$$

$$\frac{dY}{dt} = f(X_1, X_2) X_1 Y + \varepsilon_Y(X_1) \mu_Y(X_1) \rho f(X_1, X_2) X_2 Y - \delta Y$$

Throughout this paper, I use a Monod-type term to indicate resource limitation of the prey

(2) 
$$u(R) = \frac{R}{K+R}$$

and a Holling Type-2 functional response for multiple prey species to describe the predation rate by an individual predator:

(3) 
$$f(X_1, X_2) = g \frac{1}{H + X_1 + p X_2}$$



**Figure 1** – **Food web structure.** The model equations (Eq. 1) describe a diamond-shaped food web module. Solid arrows depict flows of matter due to resource or prey uptake. Dashed arrows show the interaction modification  $\mu_i(X_1)$  of the uptake rates that are caused by the provision of essential resources by the focal prey. The other potential higher-order interaction from essential resource provision  $\varepsilon_i(X_1)$  affects the conversion efficiency of the competitor or the predator and is depicted by dotted arrows. Only one of these higher-order inaction is investigated at a time.

I introduce the higher-order interactions due to essential resource provision as interaction modifications  $\mu_i(X_1)$  and  $\varepsilon_i(X_1)$  driven by the density of the focal prey species (Arditi, Michalski, and Hirzel, 2005). For generality, I include all possible options where these modifications affect the uptake rates of abiotic resources or prey, or the efficiency at which new biomass is produced, respectively. Accordingly,  $\mu_X(X_1)$  defines how an increasing density of the focal prey increases the resource uptake rate of the competing prey and  $\varepsilon_X(X_1)$  gives the conversion efficiency of those resources into new competitor biomass depending on the density of the focal prey. The same logic translates to  $\mu_Y(X_1)$  and  $\varepsilon_Y(X_1)$  for the predator, i.e. the provision of essential resources may increase the predation rate on the competitor, for example by alleviating a predator dispreference for the competitor. Similarly, the predator conversion efficiency for competitor biomass may increase due to the provision of essential resources by the focal prey. Note that in this model, I am investigating only the provision of essential resources, thus assuming that the focal prey itself always contains the essential resources. Predator consumption and conversion of focal prey biomass is thus kept constant. I assume that the modification functions  $\mu_i(X_1)$ and  $\varepsilon_i(X_1)$  monotonically increase with focal prey density, eventually approaching unity for large focal prey densities, as here their effect should vanish, as the essential resource should be abundantly present and thus non-limiting.

$$\lim_{X_1 \to \infty} \mu_i(X_1) = 1$$
$$\lim_{X_1 \to \infty} \varepsilon_i(X_1) = 1$$

The interaction modifications pose an implicit way of representing the temporal dynamics of production, distribution, stability, uptake and usage of the essential resource molecules and thus avoid the difficulties involved in modelling these processes explicitly, but explicit approaches also exist (Hammarlund, Chacón, and Harcombe, 2019; Sun et al., 2019).

I define essentiality *e* as the relative reduction of uptake rates or conversion efficiencies in the absence of the focal prey compared to when it's abundantly present and neither the uptake rates nor the conversion efficiencies are reduced. For the uptake rate modifications  $\mu_i(X_1)$  this results in

(4)  
$$e = \frac{\lim_{X_1 \to \infty} \mu_i(X_1) - \mu_i(0)}{\lim_{X_1 \to \infty} \mu_i(X_1)} = 1 - \mu_i(0)$$

**Table 1** – **Reference parameter set.** Resource concentrations and organism abundances or densities are assumed to be normalized appropriately such that their units become unity. Deviations from the reference parameter values are reported where applicable. For an illustration of the different food web scenarios see Fig. 2.

Parameter	Biological meaning	Value			
R <sub>0</sub>	Input concentration of abiotic resource	1			
δ	Chemostat dilution rate	1 time unit $^{-1}$			
r	Prey's uptake rate	2 time unit $^{-1}$			
K	Prey's half-saturation constant	0.1			
g	Predator's consumption rate	1.5 time unit $^{-1}$			
Н	Predator's half-saturation constant	0.1			
		Food web scenarios			
		I	П	111	IV
$\phi$	Relative competitiveness of the competitor	1.05	0.95	0.95	1.05
р	Relative vulnerability of the competitor to predation	0.8	0.8	1.2	1.2

A high essentiality thus implies a strong reduction in the uptake rates when the focal prey is absent. Similarly, if the higher-order interaction is incorporated into the conversion efficiencies I define

$$e = 1 - \varepsilon_i(0)$$

Note that for the sake of simplicity I investigate only the effect of one higher-order interaction at a time, i.e. only one interaction modification will be dependent on the focal prey density, keeping the other three constant at unity.

To determine the effect of essential resource provision on persistence of the focal prey I employ invasion analysis and study whether the focal prey can re-invade the resident community once it has gone extinct. This is ensured by a positive invasion growth rate which is defined as the average per-capita growth rate when rare (Ellner et al., 2019). Specifically, the invasion growth rate of the focal prey in my model is

(6) 
$$\left\langle \frac{1}{X_1} \frac{\mathrm{d}X_1}{\mathrm{d}t} \Big|_{X_1=0} \right\rangle = \langle r u(R) - f(0, X_2) Y - \delta \rangle$$

where the angle brackets denote the temporal average. If the resident community's attractor is a limit cycle, the temporal average can be obtained numerically from one period of such cycles (Ehrlich, Becks, and Gaedke, 2017). As at least some of the parameter combinations investigated in this paper result in limit cycles I used this numerical approach throughout and confirmed the results with the analytically computable solutions for those cases where the resident community was in a steady state. For numerically determining the invasion growth rate of the focal prey, the resident community dynamics were numerically integrated for 2000 time units until they reached their attractor. Convergence was determined visually and verified by ensuring that the absolute values of the slope of the linear regression on the predator abundance, as well as the slope of its moving variance was below  $10^{-3}$  per time unit. The period length was determined as in Raatz, van Velzen, and Gaedke (2019) by determining the average time spans between predator maxima during the last 200 time units using the FindMaximum algorithm in Mathematica. The average of the invasion growth rates for each time step during one period was computed. All computations were performed in Mathematica (Wolfram Research, Inc., 2019) and can be re-run using the provided Mathematica notebooks (Raatz, 2023). The analytical solutions are lengthy and can also be found in the notebook and the corresponding pdf exports.

For evaluating the state of the resident community as well as the invasion growth rate of the focal prey, the interaction modifications  $\mu_i(X_1)$  and  $\varepsilon_i(X_1)$  reduce to  $\mu_i(0) = \mu_i^0$  and  $\varepsilon_i(0) = \varepsilon_i^0$ , which I can relate to the essentiality *e* via Eqs. 4 and 5. Therefore, I obtain direct relationships

between the essentiality of the resource that is provided by the focal prey and its invasion growth rate.

Notably, the invasion analysis does not require a specific choice of the functional form of the interaction modifications. Only those numerical integrations where the focal prey is not set to zero require a particular definition. In those cases I use the following functions that monotonically increase and saturate at unity for large  $X_1$ .

(7)  
$$\mu_i(X_1) = \mu_i^0 + (1 - \mu_i^0) \frac{X_1}{0.5 R_0 + X_1}$$
$$\varepsilon_i(X_1) = \varepsilon_i^0 + (1 - \varepsilon_i^0) \frac{X_1}{0.5 R_0 + X_1}$$

To investigate under which conditions the provision of essential resources can ensure persistence I will focus on four food web scenarios that account for the non-trivial coexistence outcomes in the diamond-shaped food web module. In the first food web scenario the focal prey is the inferior competitor for resource R and is more vulnerable to predation than its competitor, which would imply extinction of the focal prey is again more vulnerable to predation but now the superior competitor for resource R, which allows for predator-mediated coexistence for a subset of the parameter space, but focal prey extinction otherwise (Abrams, 1999; Jones and Ellner, 2007) (Fig. 2c and d). Complementing these scenarios, I will scan the parameter space of vulnerability to predation p and resource competitiveness  $\phi$  of the competitor relative to the focal prey.



**Figure 2** – **The effect of higher-order interactions depends on the food web scenario.** In the first food web scenario (a), the focal prey  $X_1$  is more vulnerable to predation and less competitive than its competitor  $X_2$ , whereas it is more vulnerable to predation but also more competitive in the second food web scenario (b). The third (c) and fourth (d) food web scenarios are mirror images of the first and second food web scenario. Essentiality-mediated higher-order interactions that limit the growth of the competitor should favour persistence of the focal prey  $X_1$  (blue curved arrows). Essentiality should promote persistence of the focal prey in food webs that permit predator-mediated coexistence (blue curved arrow in (b)), but likely is detrimental otherwise (red curved arrows) as it can render the competitor effectively less vulnerable to predation than the focal prey.

A priori one would expect that essentiality that limits the growth and competitiveness of the competitor should favour the persistence of the focal prey. Further, I hypothesize that within predator-mediated coexistence an increasing essentiality should make the focal prey more indispensable to the community and therefore increase its invasion growth rate, possibly even

eventually fulfilling the invasion criterion

$$\left\langle \frac{1}{X_1} \frac{\mathsf{d}X_1}{\mathsf{d}t} \bigg|_{X_1=0} \right\rangle > 0$$

that would prove an ensured persistence of the focal prey.

#### Results

Scanning the parameter space of vulnerability to predation p and resource competitiveness  $\phi$  of the competitor relative to the focal prey provides an overview of the effects of essentiality on the persistence of the focal prey (Fig. 3). Comparing the invasion growth rates at vanishing and complete essentiality. I find that depending on these parameters, and thus the respective food web scenario, essentiality-mediated higher-order interactions can promote but also counter-act the persistence of the focal prey, or have no effect as the focal prey persists or goes extinct irrespective of its essentiality. Analysing the four food web scenarios in more detail provides a detailed understanding of the mechanisms behind these patterns.

In the first food web scenario (Fig. 2a) the focal prey does not persist for vanishing essentiality, as indicated by a negative invasion growth rate. However, increasing essentiality when the higher-order interaction affects the resource uptake rate or conversion efficiency of the competitor turns the invasion growth rate positive (Fig. 4a,b) and thus ensures the persistence of the focal prey (Fig. 5). This includes a drastic shift in the resident community shortly beyond e = 0.4 where first the predator and then the prey go extinct (Fig. A1a,b). An essentiality of e = 0.4 implies that the resource uptake rate or the conversion efficiency of the competitor are reduced to 60% in the absence of the focal prey. In my model formulation this implies that the competitor cannot sustain the predator further which, in the absence of the focal prey, therefore goes extinct. A slight additional reduction hinders the competitor from outgrowing dilution and thus drives it to extinction as well. In this food web scenario, higher-order interactions that target the uptake rate or the conversion efficiency of the predator do not benefit the persistence of the focal prey (Fig. 4c,d) due to unfavourable trait combinations. As the focal prey is the inferior competitor for the resource R and also more vulnerable to predation it can persist neither in the absence nor in the presence of the predator. Supporting the predator by providing essential resources harms the focal prey more than the competitor. For the predator, a larger dependence on the focal prey is also disadvantageous as this decreases its uptake rate and conversion efficiency, and results in extinction at approximately e = 0.25 (Fig. A1c,d).

In the second food web scenario, the focal prey is still more vulnerable to predation than its competitor but now it is also the superior competitor for the resource R (Fig. 3). While the invasion analysis outcomes are similar to the first food web scenario for higher-order interactions targeting the competitor's uptake rate or conversion efficiency (Fig. 6a,b), the trait combinations now allow for positive invasion growth rates also when the higher-order interaction targets the predator's uptake rate or conversion efficiency (Fig. 6c,d). Therefore, increasing essentiality can promote the persistence of the focal prey for intermediate to large essentiality in this food web scenario. This persistence-promoting effect of essentiality appears in a parameter range of predator-mediated coexistence of prey (Fig. 7). Here, the predator goes extinct in the resident community as the competing prey alone does not sustain the predator given the reduction in uptake rate or conversion efficiency for large essentiality of the focal prey (Fig. A2). In the absence of the predator the focal prey benefits from its higher competitiveness for the resource R and thus persists. Once it invades it may additionally sustain the predator (Fig. 7b). Conditional on the presence or absence of the predator when the focal prey invades two community states are therefore possible. Without the predator the focal prey out-competes the competitor which thus goes extinct (Fig. 7b, solid lines). If the predator is present or is re-introduced it however mediates coexistence of the focal prey and the competitor (Fig. 7b, dashed lines). This shows that providing essential resources can affect not only the focal prey itself, but also the whole community structure.





In the third food web scenario, the focal prey persists independent of essentiality as indicated by a positive invasion growth rate for all possible types of essentiality-mediated higher order interactions (Fig. 3 and Fig. A3). If essentiality affects the competitor the focal prey's invasion growth rate increases further. If, however, essentiality causes limitations for the predator the invasion growth rate tends to decrease for larger essentiality (albeit not turning negative) as this effectively reduces the energy flow from the competitor to the predator and thus eventually renders the competitor less vulnerable to predation than the focal prey.

Similarly, higher-order interactions affecting the competitor increase the invasion growth rate of the focal prey with higher essentiality in the fourth food web scenario. If the resource competitiveness of the competitor is only slightly exceeding the resource competitiveness of the



**Figure 4** – **Invasion growth rate of the focal prey for the first food web scenario (Fig. 2a).** Here, the focal prey is more vulnerable to predation and competitively inferior to the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The grey shading indicates the states of the resident community (Fig. A1). For dark shading both predator and competitor coexist, for light-grey shading only the competitor persists and for no shading only the resource remains. The analytically computed invasion growth rate (dashed line) deviates from the numerical observation (full line) once the dynamics become cyclic. The vertical dotted line marks the bifurcation point.



Figure 5 – Population dynamics for the first food web scenario when the higher-order interaction targets the resource uptake rate of the competitor (Fig. 4a). (a) For small essentialities the invasion rate of the focal prey is negative and it thus cannot invade. (b) A larger essentiality ensures the persistence of the focal prey. To obtain these dynamics, I chose the interaction modification according to Eq. 7, integrated the resident community to its stable state and then introduced the focal prey at an initial biomass of  $X_{1,0} = 10^{-3}$ .

focal prey the invasion growth rate of the focal prey is positive even for zero essentiality and



**Figure 6** – **Invasion growth rate for the second food web scenario.** Here, the focal prey is more vulnerable to predation but also competitively superior to the competitor. In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The plot specifics are identical to Fig. 4.



Figure 7 – Population dynamics for the second food web scenario when the higherorder interaction targets the resource uptake rate of the shared predator (Fig. 6c). (a) For small essentialities the invasion rate of the focal prey is negative and it thus cannot invade. (b) A larger essentiality ensures the persistence of the focal prey. To obtain these dynamics, I chose the interaction modification according to Eq. 7, integrated the resident community to its stable state and then introduced the focal prey at an initial biomass of  $X_{1,0} = 10^{-3}$ . In panel (b) the predator goes extinct in the residence community, thus I assumed  $Y_0 = 0$  (thick lines). If, however, the predator is reintroduced together with the focal prey ( $Y_0 = 10^{-3}$ , thin dashed lines), it is supported by the focal prey, re-establishes and mediates the coexistence of both prey types.

only increases further for higher essentiality (Fig. A4). For higher competitiveness of the competitor the invasion growth rate at zero essentiality is negative and turns positive for higher

essentiality, again resulting in promoted persistence already (Fig. 3a,b). In this scenario, however, higher-order interactions affecting the predator can result in a negative invasion growth rate, which can become even smaller if the essentiality becomes larger. Here again, an increasing essentiality counteracts the larger vulnerability of the competitor to predation, and allows the competitor to outcompete the focal prey given its higher resource competitiveness.

#### Discussion

Higher-order interactions have the potential to shape community structure and dynamics (Grilli et al., 2017; Mayfield and Stouffer, 2017; Terry, Morris, and Bonsall, 2019). In this paper, I showed how the provision of essential resources creates a higher-order interaction that decisively affects the persistence of the focal prey and the resulting community structure. I investigated both the case of essential resource provision to community members from the same trophic level as well as from a higher trophic level. Whether these higher-order interactions in the end ensure persistence depends both on their strength as well as on the food web scenario (see Fig. 2 for a summary of the results).

Confirming the expectations, I find in all food web scenarios that a larger essentiality for the competitor can increase the invasion growth rate of the focal prey. In the first and second food web scenario where the invasion growth rate is negative for zero essentiality this leads to a sign-change in the invasion growth rate and thus a promoting effect of essentiality on persistence. In the third and fourth food web scenarios the invasion growth rate of the focal prey is already positive for zero essentiality and only increases further for larger essentiality. Essentiality for the predator can indeed favour the persistence of the focal prey in food web scenarios that permit predator-mediated coexistence of the prey species (second food web scenario), but can also be detrimental for persistence if it renders the competitor effectively less vulnerable to predation (fourth food web scenario). Further, I find that essentiality determines the resident community structure, with larger essentiality driving extinct first the predator and then, depending on the higher-order interaction, potentially also the competitor. As seen in the second food web scenario this allows for multiple possible community states, depending on whether the coexistence-mediating predator is re-introduced together with the focal prey. Further, no qualitative differences between higher-order interactions affecting the uptake rate or the conversion efficiency were observed.

Experimental support exists for both higher-order interactions that affect the uptake rate or the conversion efficiency. Essential resources affecting the uptake rate could result from adaptive foraging behaviour, as predicted by nutritional geometry (Raubenheimer and Simpson, 1993; Simpson et al., 2004), selective feeding (Buskey, 1997; Eberl et al., 2020; Elser et al., 2016; Meunier et al., 2016), or changed behaviour due to the provision of essential micronutrients, as recently reported for a nematode feeding on larvae of other nematodes (Akduman et al., 2020). Here, the attack rate of the predatory nematode increased when reared on vitamin  $B_{12}$  producing bacteria compared to  $B_{12}$  deficient controls. However, feeding rate was not increased in this study, so only the prey's loss term would be affected by this higher-order interaction when transferring these results to my model. Another possibility would be generally better physiological conditions that increase fitness, as reported for *Daphnia magna* and vitamin  $B_{12}$  (Kusari et al., 2017), which could also translate to generally increased activity.

The most direct and intuitive mechanism for a higher-order interaction that affects the conversion efficiency of a consumer via essential resource provision is that those lacking essential nutrients that are halting biomass production are directly provided. This is the case in the above example with *Daphnia magna* and vitamin B<sub>12</sub> (Keating, 1985), other nutrients like phosphorous (Urabe, Shimizu, and Yamaguchi, 2018) or biochemicals (Eberl et al., 2020; Raatz, Gaedke, and Wacker, 2017). Similarly, supplementing herbivory with fungivory was found to significantly speed up growth in moth larvae (Eberl et al., 2020). Microbial cross-feeding likely represents the case of higher-order interactions affecting the conversion efficiency of organisms on the same trophic level (D'Souza et al., 2018). In the absence of another carbon source bacteria depend on algal carbon fixation and exudation (Bratbak and Thingstad, 1985; Raatz, Schälicke, et al., 2018),

which was proposed as the mutualistic trade in return for bacterial vitamin  $B_{12}$  provision (Croft et al., 2005) during this type of cross-feeding between different kingdoms.

There has been a long history of investigating the effect of higher-order interactions in small ecological interaction networks, such as trait-mediated indirect interactions (Bolker et al., 2003; Werner and Peacor, 2003) or non-trophic interactions (Kéfi, Berlow, Wieters, Navarrete, et al., 2012), e.g. facilitation (Gross, 2008). The effect of higher-order interactions on community stability is investigated also in larger networks, both empirical (González González et al., 2021) and theoretical, randomly sampled ones e.g. Arditi, Michalski, and Hirzel, 2005; Gibbs et al., 2023; Grilli et al., 2017, and innovative approaches of analyzing their effects have been proposed (Golubski et al., 2016). The effect of trait-mediated indirect interactions and higher-order interactions in general have been shown to depend on many specifics, such as network structure and interaction strengths. In my model, a higher essentiality corresponds to a higher strength of the higher-order interaction. I found that depending on the food web scenario, food-quality-provision-mediated higher-order interactions can be both promoting but also detrimental to persistence and thus community stability, a finding that resonates with this overall complexity. Exploring the effect of multiple, simultaneously occurring higher-order interactions presents an interesting avenue for future research.

The provision of essential resources changes the abiotic environment of the competitors or predators via changing the pool of available essential resources. It can be seen as a form of niche construction that is implicitly included via an interaction modification between two biotic food web components similar to Kylafis and Loreau, 2011; Oña, Giri, et al., 2021. Obviously, the niches of predator and competitor are impacted directly by the presence of the focal prey. Interestingly, however, this niche construction operates also indirectly in the second food web scenario, as the niche of the focal prey is extended through a feedback loop via predator-mediated coexistence of competitor and focal prey.

Bridging theory and experiments on higher-order-interactions is challenging (Levine et al., 2017). I worked out that essentiality, defined as the reduction of uptake rates or conversion efficiencies when the focal prey is absent, is an appropriate measure to determine the effect of the higher-order interaction on the persistence of the focal prey, particularly when using invasion analysis. One of the benefits from this definition is that the density-dependent functional form of the higher-order interactions does not need to be specified, which largely facilitates experimental approaches of measuring the presence and effect of the higher-order interactions. In my analysis I focussed on the persistence of the focal prey. It should be noted that determining coexistence of species, and not only persistence of a focal species, can be complicated by the existence of multiple stable states e.g. Yamamichi, Yoshida, and Sasaki, 2014 which constrains the interpretation of invasion growth rates (Grainger, Levine, and Gilbert, 2019).

Measuring higher-order interactions experimentally is difficult, however, some advances have been reported that employ different strategies. A first line of research infers the higher-order interactions statistically from community dynamics data (e.g. Kéfi, Berlow, Wieters, Joppa, et al., 2015; Mayfield and Stouffer, 2017). A second, more mechanistic approach aims to disable hypothesized higher-order interactions and compare the outcomes with the non-manipulated scenario. One prominent example is the study by Wootton (1993) where the disguising effect of barnacles for limpets was discovered by removing barnacles partially or completely. Removing the species that initiates the higher-order interaction to quantify the effect of the higherorder interaction however is complicated by other direct and indirect effects that are then also removed, which would lead to false evaluations of the effect size of the higher-order interaction. The essential resource context provides a different way of determining the effect size of higher-order interactions. Experimentally providing the essential resource in excess by supplementation removes its potential to cause higher-order interactions and decreases its essentiality. This approach has been used in investigations of microbial cross-feeding, such as in Kazamia et al. (2012) and Hammarlund, Chacón, and Harcombe (2019) where supplementation with the essential resource alleviated the dependence on the interaction partner, shifting the coexistence pattern towards the beneficiary of the supplementation. In the context of herbivore limitation by biochemicals, supplementation was used to show the mechanistic basis for the higher-order

interaction (Wacker and Martin-Creuzburg, 2012). In a predator-prey context Bayesian inference from population size time series can be used to obtain uptake rates and conversion efficiencies independently (Rosenbaum et al., 2019). Applying the inference for different supplementation levels should allow to disentangle whether the essential resource affects the uptake rate or the conversion efficiency of the predator. This may be less feasible for a prey consuming abiotic resources, but even here methods such as isotopic labelling could be used to track uptake and conversion separately. The community-structuring effect of essential resource provision remains to be tested, which requires tracking the population feedback mechanisms over larger time scales of many prey generations, but chemostat or mesocosm experiments will be useful here. The central focus of this article on persistence of the focal prey, however, facilitates experimental validation. As argued before, only the invasion growth rate of the focal prey would have to be obtained for different levels of supplementation with potentially different resident communities. This reduces the time that experimental cultures would have to be operated and avoids experimental difficulties often entailed by long-term observations, ultimately illuminating the potential effect of essential resource provision on prey persistence.

# Acknowledgements

The author would like to thank Elias Ehrlich for helpful comments on an earlier draft, as well as two reviewers and the recommender Cédric Gaucherel for their constructive suggestions to improve this manuscript. Preprint version 3 of this article has been peer-reviewed and recommended by Peer Community In Ecology (Gaucherel, 2023, https://doi.org/10.24072/pci.ecology.100527).

# Fundings

Funding for this research was provided by Max Planck Society.

# Conflict of interest disclosure

The author declares that he complies with the PCI rule of having no financial conflicts of interest in relation to the content of the article.

# Data, script, code, and supplementary information availability

The Mathematica scripts used for the calculations and for creating the figures have been deposited both as Mathematica notebooks and also as pdf exports in a Zenodo repository (Raatz, 2023, https://doi.org/10.5281/zenodo.7575588).

# References

- Abrams PA (1999). Is Predator-Mediated Coexistence Possible in Unstable Systems? Ecology 80, 608-621. https://doi.org/10.2307/176639.
- Adamowicz EM, Flynn J, Hunter RC, Harcombe WR (2018). Cross-Feeding Modulates Antibiotic Tolerance in Bacterial Communities. The ISME Journal **12** (11), 2723–2735. https://doi.org/10.1038/s41396-018-0212-z.
- Akduman N, Lightfoot JW, Röseler W, Witte H, Lo WS, Rödelsperger C, Sommer RJ (2020). Bacterial Vitamin B12 Production Enhances Nematode Predatory Behavior. The ISME Journal 14, 1494–1507. https://doi.org/10.1038/s41396-020-0626-2.
- Andersen T, Elser JJ, Hessen DO (2004). Stoichiometry and Population Dynamics: Stoichiometry and Population Dynamics. Ecology Letters 7, 884–900. https://doi.org/10.1111/j.1461–0248.2004.00646.x.
- Anderson TR, Hessen DO (2005). Threshold Elemental Ratios for Carbon versus Phosphorus Limitation in Daphnia. Freshwater Biology **50**, 2063–2075. https://doi.org/10.1111/j.1365– 2427.2005.01450.x.
- Arditi R, Michalski J, Hirzel AH (2005). Rheagogies: Modelling Non-Trophic Effects in Food Webs. Ecological Complexity 2, 249–258. https://doi.org/10.1016/j.ecocom.2005.04.003.

- Azam F, Fenchel T, Field J, Gray J, Meyer-Reil L, Thingstad F (1983). The Ecological Role of Water-Column Microbes in the Sea. Marine Ecology Progress Series 10 (November 2015), 257–263. https://doi.org/10.3354/meps010257.
- Billick I, Case TJ (1994). Higher Order Interactions in Ecological Communities: What Are They and How Can They Be Detected? Ecology **75**, 1529–1543. https://doi.org/10.2307/1939614.
- Bolker B, Holyoak M, Křivan V, Rowe L, Schmitz O (2003). *Connecting Theoretical and Empirical Studies of Trait-Mediated Interactions*. *Ecology* **84**, 1101–1114. https://doi.org/10.1890/0012-9658(2003)084[1101:CTAES0]2.0.C0;2.
- Bratbak G, Thingstad TF (1985). Phytoplankton-Bacteria Interactions: An Apparent Paradox? Analysis of a Model System with Both Competition and Commensalism. Marine Ecology Progress Series **25**, 23–30. https://doi.org/10.1016/0198–0254(86)91170–2.
- Burian A, Nielsen JM, Winder M (2020). Food Quantity-Quality Interactions and Their Impact on Consumer Behavior and Trophic Transfer. Ecological Monographs **90**. https://doi.org/10.1002/ecm.1395.
- Buskey E (1997). Behavioral Components of Feeding Selectivity of the Heterotrophic Dinoflagellate Protoperidinium Pellucidum. Marine Ecology Progress Series **153**, 77–89. https://doi.org/10. 3354/meps153077.
- Chesson P (1994). Multispecies Competition in Variable Environments. Theoretical Population Biology **45**, 227–276. https://doi.org/10.1006/tpbi.1994.1013.
- Croft MT, Lawrence AD, Raux-Deery E, Warren MJ, Smith AG (2005). Algae Acquire Vitamin B12 through a Symbiotic Relationship with Bacteria. Nature **438**, 90–93. https://doi.org/10.1038/nature04056.
- D'Souza G, Shitut S, Preussger D, Yousif G, Waschina S, Kost C (2018). Ecology and Evolution of Metabolic Cross-Feeding Interactions in Bacteria. Natural Product Reports **35**, 455–488. https://doi.org/10.1039/c8np00009c.
- Douglas AE (2015). Multiorganismal Insects: Diversity and Function of Resident Microorganisms. Annual Review of Entomology **60**, 17–34. https://doi.org/10.1146/annurev-ento-010814-020822.
- Eberl F, Fernandez de Bobadilla M, Reichelt M, Hammerbacher A, Gershenzon J, Unsicker SB (2020). Herbivory Meets Fungivory: Insect Herbivores Feed on Plant Pathogenic Fungi for Their Own Benefit. Ecology Letters 23. Ed. by Ted Turlings, 1073–1084. https://doi.org/10.1111/ele.13506.
- Ehrlich E, Becks L, Gaedke U (2017). Trait-Fitness Relationships Determine How Trade-off Shapes Affect Species Coexistence. Ecology **98**, 3188–3198. https://doi.org/10.1002/ecy.2047.
- Ehrlich E, Gaedke U (2018). Not Attackable or Not Crackable How Pre- and Post-Attack Defenses with Different Competition Costs Affect Prey Coexistence and Population Dynamics. Ecology and Evolution 8, 6625–6637. https://doi.org/10.1002/ece3.4145.
- Ellner SP, Snyder RE, Adler PB, Hooker G (2019). An Expanded Modern Coexistence Theory for Empirical Applications. Ecology Letters 22. Ed. by Jessica Metcalf, 3–18. https://doi.org/10.1111/ele.13159.
- Elser J, Kyle M, Learned J, McCrackin M, Peace A, Steger L (2016). Life on the Stoichiometric Knife-Edge: Effects of High and Low Food C:P Ratio on Growth, Feeding, and Respiration in Three Daphnia Species. Inland Waters 6, 136–146. https://doi.org/10.5268/IW-6.2.908.
- Filipiak M, Kuszewska K, Asselman M, Denisow B, Stawiarz E, Woyciechowski M, Weiner J (2017). Ecological Stoichiometry of the Honeybee: Pollen Diversity and Adequate Species Composition Are Needed to Mitigate Limitations Imposed on the Growth and Development of Bees by Pollen Quality. PLOS ONE 12, e0183236. https://doi.org/10.1371/journal.pone.0183236.
- Gaedke U, Hochstädter S, Straile D (2002). Interplay between Energy Limitation and Nutritional Deficiency: Empirical Data and Food Web Models. Ecological Monographs **72**, 251–270. https://doi.org/10.1890/0012-9615(2002)072[0251:IBELAN]2.0.C0;2.
- Gaucherel C (2023). High-Order Interactions in Food Webs May Strongly Impact Persistence of Species. Peer Community in Ecology. https://doi.org/10.24072/pci.ecology.100527.

14

- Gibbs T, Gellner G, Levin SA, McCann KS, Hastings A, Levine JM (2023). Can Higher-Order Interactions Resolve the Species Coexistence Paradox? bioRxiv. https://doi.org/10.1101/2023. 06.19.545649.
- Golubski AJ, Westlund EE, Vandermeer J, Pascual M (2016). *Ecological Networks over the Edge: Hypergraph Trait-Mediated Indirect Interaction (TMII) Structure. Trends in Ecology & Evolution* **31**, 344–354. https://doi.org/10.1016/j.tree.2016.02.006.
- González González C, Mora Van Cauwelaert E, Boyer D, Perfecto I, Vandermeer J, Benítez M (2021). High-Order Interactions Maintain or Enhance Structural Robustness of a Coffee Agroecosystem Network. Ecological Complexity **47**, 100951. https://doi.org/10.1016/j.ecocom. 2021.100951.
- Gore J, Youk H, Van Oudenaarden A (2009). Snowdrift Game Dynamics and Facultative Cheating in Yeast. Nature **459**, 253–256. https://doi.org/10.1038/nature07921.
- Grainger TN, Levine JM, Gilbert B (2019). The Invasion Criterion: A Common Currency for Ecological Research. Trends in Ecology & Evolution **34**, 925–935. https://doi.org/10.1016/j.tree. 2019.05.007.
- Grilli J, Barabás G, Michalska-Smith MJ, Allesina S (2017). *Higher-Order Interactions Stabilize Dynamics in Competitive Network Models*. *Nature* **548**, 210–213. https://doi.org/10.1038/ nature23273.
- Gross K (2008). Positive Interactions among Competitors Can Produce Species-Rich Communities. Ecology Letters **11**, 929–936. https://doi.org/10.1111/j.1461-0248.2008.01204.x.
- Guo F, Kainz MJ, Sheldon F, Bunn SE (2016). The Importance of High-Quality Algal Food Sources in Stream Food Webs - Current Status and Future Perspectives. Freshwater Biology **61**, 815–831. https://doi.org/10.1111/fwb.12755.
- Hammarlund SP, Chacón JM, Harcombe WR (2019). A Shared Limiting Resource Leads to Competitive Exclusion in a Cross-Feeding System. Environmental Microbiology **21**, 759–771. https: //doi.org/10.1111/1462-2920.14493.
- Herren CM (2020). Disruption of Cross-Feeding Interactions by Invading Taxa Can Cause Invasional Meltdown in Microbial Communities. Proceedings of the Royal Society B: Biological Sciences **287**, 20192945. https://doi.org/10.1098/rspb.2019.2945.
- Iwabuchi T, Urabe J (2012). Food Quality and Food Threshold: Implications of Food Stoichiometry to Competitive Ability of Herbivore Plankton. Ecosphere 3, art51-art51. https://doi.org/10. 1890/ES12-00098.1.
- Jones LE, Ellner SP (2007). Effects of Rapid Prey Evolution on Predator-Prey Cycles. Journal of Mathematical Biology **55**, 541–573. https://doi.org/10.1007/s00285-007-0094-6.
- Kazamia E, Czesnick H, Nguyen TTV, Croft MT, Sherwood E, Sasso S, Hodson SJ, Warren MJ, Smith AG (2012). Mutualistic Interactions between Vitamin B12-dependent Algae and Heterotrophic Bacteria Exhibit Regulation. Environmental Microbiology 14, 1466–1476. https://doi.org/ 10.1111/j.1462-2920.2012.02733.x.
- Keating KI (1985). The Influence of Vitamin B12 Deficiency on the Reproduction of Daphnia Pulex Leydig (Cladocera). Journal of Crustacean Biology 5, 130–136. https://doi.org/10.2307/1548225.
- Kéfi S, Berlow EL, Wieters EA, Joppa LN, Wood SA, Brose U, Navarrete SA (2015). *Network Structure beyond Food Webs: Mapping Non-Trophic and Trophic Interactions on Chilean Rocky Shores. Ecology* **96**, 291–303. https://doi.org/10.1890/13-1424.1.
- Kéfi S, Berlow EL, Wieters EA, Navarrete SA, et al. (2012). More than a Meal... Integrating Non-Feeding Interactions into Food Webs. Ecology Letters **15**, 291–300. https://doi.org/10. 1111/j.1461-0248.2011.01732.x.
- Koedooder C, Stock W, Willems A, Mangelinckx S, De Troch M, Vyverman W, Sabbe K (2019). Diatom-Bacteria Interactions Modulate the Composition and Productivity of Benthic Diatom Biofilms. Frontiers in Microbiology **10**, 1255. https://doi.org/10.3389/fmicb.2019.01255.
- Koussoroplis AM, Schälicke S, Raatz M, Bach M, Wacker A (2019). Feeding in the Frequency Domain: Coarser-Grained Environments Increase Consumer Sensitivity to Resource Variability, Covariance and Phase. Ecology Letters **22**, 1104–1114. https://doi.org/10.1111/ele.13267.

- Kusari F, O'Doherty AM, Hodges NJ, Wojewodzic MW (2017). Bi-Directional Effects of Vitamin B 12 and Methotrexate on Daphnia Magna Fitness and Genomic Methylation. Scientific Reports 7 (1), 11872. https://doi.org/10.1038/s41598-017-12148-2.
- Kylafis G, Loreau M (2008). Ecological and Evolutionary Consequences of Niche Construction for Its Agent. Ecology Letters **11**, 1072–1081. https://doi.org/10.1111/j.1461-0248.2008. 01220.x.
- Kylafis G, Loreau M (2011). Niche Construction in the Light of Niche Theory. Ecology Letters 14, 82–90. https://doi.org/10.1111/j.1461-0248.2010.01551.x.
- Laland K, Matthews B, Feldman MW (2016). An Introduction to Niche Construction Theory. Evolutionary Ecology **30**, 191–202. https://doi.org/10.1007/s10682-016-9821-z.
- Levine JM, Bascompte J, Adler PB, Allesina S (2017). Beyond Pairwise Mechanisms of Species Coexistence in Complex Communities. Nature 546 (7656), 56–64. https://doi.org/10.1038/ nature22898.
- MacArthur R, Levins R (1967). The Limiting Similarity, Convergence, and Divergence of Coexisting Species. The American Naturalist **101**, 377–385. https://doi.org/10.1086/282505.
- Mayfield MM, Stouffer DB (2017). Higher-Order Interactions Capture Unexplained Complexity in Diverse Communities. Nature Ecology and Evolution 1, 1–7. https://doi.org/10.1038/s41559-016-0062.
- Meunier CL, Boersma M, Wiltshire KH, Malzahn AM (2016). Zooplankton Eat What They Need: Copepod Selective Feeding and Potential Consequences for Marine Systems. Oikos **125**, 50–58. https://doi.org/10.1111/oik.02072.
- Miele V, Guill C, Ramos-Jiliberto R, Kéfi S (2019). Non-Trophic Interactions Strengthen the Diversity—Functioning Relationship in an Ecological Bioenergetic Network Model. PLoS Computational Biology 15. https://doi.org/10.1371/journal.pcbi.1007269. pmid: 31465440.
- Muller EB, Nisbet RM, Kooijman SALM, Elser JJ, McCauley E (2001). Stoichiometric Food Quality and Herbivore Dynamics. Ecology Letters 4, 519–529. https://doi.org/10.1046/j.1461–0248.2001.00240.x.
- Oña L, Giri S, Avermann N, Kreienbaum M, Thormann KM, Kost C (2021). Obligate Cross-Feeding Expands the Metabolic Niche of Bacteria. Nature Ecology & Evolution **5** (9), 1224–1232. https://doi.org/10.1038/s41559-021-01505-0.
- Oña L, Kost C (2022). Cooperation Increases Robustness to Ecological Disturbance in Microbial Cross-Feeding Networks. Ecology Letters **25**, 1410–1420. https://doi.org/10.1111/ele.14006.
- Pomeroy L, Williams PJ, Azam F, Hobbie J (2007). The Microbial Loop. Oceanography 20, 28–33. https://doi.org/10.5670/oceanog.2007.45.
- Raatz M, Schälicke S, Sieber M, Wacker A, Gaedke U (2018). One Man's Trash Is Another Man's Treasure-the Effect of Bacteria on Phytoplankton-Zooplankton Interactions in Chemostat Systems: Bacteria in Chemostat Experiments. Limnology and Oceanography: Methods **16**, 629–639. https://doi.org/10.1002/lom3.10269.
- Raatz M (2023). Provision of Essential Resources as a Persistence Strategy in Food Webs Code. Zenodo. https://doi.org/10.5281/zenodo.8269141.
- Raatz M, Gaedke U, Wacker A (2017). High Food Quality of Prey Lowers Its Risk of Extinction. Oikos **126**, 1501–1510. https://doi.org/10.1111/oik.03863.
- Raatz M, Velzen E, Gaedke U (2019). Co-Adaptation Impacts the Robustness of Predator-Prey Dynamics against Perturbations. Ecology and Evolution 9, 3823–3836. https://doi.org/10. 1002/ece3.5006.
- Raubenheimer D, Simpson SJ (1993). The Geometry of Compensatory Feeding in the Locust. Animal Behaviour **45**, 953–964. https://doi.org/10.1006/anbe.1993.1114.
- Rosenbaum B, Raatz M, Weithoff G, Fussmann GF, Gaedke U (2019). *Estimating Parameters From Multiple Time Series of Population Dynamics Using Bayesian Inference. Frontiers in Ecology and Evolution* **6**, 234–234. https://doi.org/10.3389/fevo.2018.00234.
- Schade JD, Kyle M, Hobbie SE, Fagan WF, Elser JJ (2003). Stoichiometric Tracking of Soil Nutrients by a Desert Insect Herbivore. Ecology Letters 6, 96–101. https://doi.org/10.1046/j.1461-0248.2003.00409.x.

- Schälicke S, Sobisch LY, Martin-Creuzburg D, Wacker A (2019). Food Quantity–Quality Co-Limitation: Interactive Effects of Dietary Carbon and Essential Lipid Supply on Population Growth of a Freshwater Rotifer. Freshwater Biology **64**, 903–912. https://doi.org/10.1111/fwb.13272.
- Simpson SJ, Sibly RM, Lee KP, Behmer ST, Raubenheimer D (2004). Optimal Foraging When Regulating Intake of Multiple Nutrients. Animal Behaviour 68, 1299–1311. https://doi.org/10. 1016/j.anbehav.2004.03.003.
- Singer MS, Farkas TE, Skorik CM, Mooney K (2012). Tritrophic Interactions at a Community Level: Effects of Host Plant Species Quality on Bird Predation of Caterpillars. The American naturalist **179**, 363–374. https://doi.org/10.1086/664080.
- Soria-Dengg S, Reissbrodt R, Horstmann U (2001). Siderophores in Marine Coastal Waters and Their Relevance for Iron Uptake by Phytoplankton: Experiments with the Diatom Phaeodactylum Tricornutum. Marine Ecology Progress Series **220**, 73–82. https://doi.org/10.3354/ meps220073.
- Stiefs D, Voorn GAK, Kooi BW, Feudel U, Gross T (2010). Food Quality in Producer-Grazer Models: A Generalized Analysis. The American Naturalist **176**, 367–380. https://doi.org/10.1086/ 655429.
- Suleiman M, Zecher K, Yücel O, Jagmann N, Philipp B (2016). Interkingdom Cross-Feeding of Ammonium from Marine Methylamine-Degrading Bacteria to the Diatom Phaeodactylum Tricornutum. Applied and Environmental Microbiology 82. Ed. by A. J. M. Stams, 7113–7122. https://doi.org/10.1128/AEM.01642–16.
- Sun Z, Koffel T, Stump SM, Grimaud GM, Klausmeier CA (2019). Microbial Cross-Feeding Promotes Multiple Stable States and Species Coexistence, but Also Susceptibility to Cheaters. Journal of Theoretical Biology **465**, 63–77. https://doi.org/10.1016/j.jtbi.2019.01.009.
- Terry JCD, Morris RJ, Bonsall MB (2019). Interaction Modifications Lead to Greater Robustness than Pairwise Non-Trophic Effects in Food Webs. Journal of Animal Ecology **88**, 1732–1742. https://doi.org/10.1111/1365-2656.13057.
- Urabe J, Shimizu Y, Yamaguchi T (2018). Understanding the Stoichiometric Limitation of Herbivore Growth: The Importance of Feeding and Assimilation Flexibilities. Ecology Letters **21**. Ed. by Robert Sterner, 197–206. https://doi.org/10.1111/ele.12882.
- Wacker A, Martin-Creuzburg D (2012). Biochemical Nutrient Requirements of the Rotifer Brachionus Calyciflorus: Co-limitation by Sterols and Amino Acids. Functional Ecology **26**, 1135–1143. https: //doi.org/10.1111/j.1365-2435.2012.02047.x.
- Werner EE, Peacor SD (2003). A Review of Trait-mediated Indirect Interactions in Ecological Communities. Ecology 84, 1083-1100. https://doi.org/10.1890/0012-9658(2003)084[1083: AROTII]2.0.C0;2.
- Wolfram Research, Inc. (2019). *Mathematica*, Version 12.0. Wolfram Research, Inc., Champaign, IL.
- Wootton JT (1993). Indirect Effects and Habitat Use in an Intertidal Community: Interaction Chains and Interaction Modifications. The American Naturalist **141**, 71–89. https://doi.org/10.1086/285461.
- Wootton JT (2002). Indirect Effects in Complex Ecosystems: Recent Progress and Future Challenges. Journal of Sea Research 48, 157–172. https://doi.org/10.1016/S1385–1101(02)00149–1.
- Yamamichi M, Yoshida T, Sasaki A (2014). Timing and Propagule Size of Invasion Determine Its Success by a Time-Varying Threshold of Demographic Regime Shift. Ecology **95**, 2303–2315. https://doi.org/10.1890/13-1527.1.

# **Appendix - Supporting figures**



**Figure A1** – **Resident community states for the first food web scenario.** In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator, which affects their community dynamics. Full lines represent the minima and maxima of one population cycle, if the population is cycling, or otherwise the steady state biomass. The vertical dotted line indicates the bifurcation point. Population dynamics were defined as cyclic if the difference between predator extrema exceeded  $10^{-5}$ . During cycles, the unstable fixed point is indicated by the dashed line. As in Fig. 4, the grey shading indicates the states of the resident community. For dark shading both predator and competitor coexist, for light-grey shading only the competitor persists and for no shading only the resource remains.



**Figure A2** – **Resident community states for the second food web scenario.** Here, the focal prey is more vulnerable to predation but also the superior competitor for the abiotic resource. In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The plot specifics are identical to Fig. A1.



**Figure A3** – **Invasion growth rate of the focal prey for the third food web scenario (Fig. 2c).** Here, the focal prey is less vulnerable to predation and more competitive for the resource than the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. Further plot specifics are identical to Figs. 4 and 6.



**Figure A4** – **Invasion growth rate of the focal prey for the fourth food web scenario (Fig. 2d).** Here, the focal prey is less vulnerable to predation and less competitive for the resource than the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. Further plot specifics are identical to Figs. 4 and 6.